

Mestrado em Biologia e Ecologia do Litoral Marinho

Conservation Priority Index for the Portuguese Estuarine Fish Species

Índice de Prioridade de Conservação para as Espécies de Peixes
Estuarinas Portuguesas

Paulo Branco

Orientador: Prof. Doutor Pedro Raposo de Almeida

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Conservation Priority Index for the Portuguese Estuarine Fish Species

Abstract

Environmental awareness is increasing and more efficient methodologies in determining the conservation priorities of biological components are needed. Multimetric indices are important tools, allowing data simplification and consequently a better understanding by managers, decision makers and general public. The aim of this work is to formulate an index capable of ranking the Portuguese estuarine fish species by their conservation priority. Ten metrics were scored, with bibliography references, for 72 species of 16 estuarine systems (W and S coasts of Portugal). Index Validation was done by external means. This index is the first to specifically address the conservation priority of estuarine fish and allows an identification of the top and bottom conservation priority species adding thus the planning of estuarine fish conservation.

Resumo

A consciência ambiental está a aumentar, sendo necessárias metodologias mais eficientes para determinar as prioridades de conservação de componentes biológicos. Os índices multimétricos são ferramentas importantes que permitem simplificar os dados, facilitando a sua compreensão por parte de gestores, decisores e do público em geral. O objectivo deste trabalho é formular um índice capaz de ordenar as espécies estuarinas portuguesas de peixes de acordo com a sua prioridade de conservação. Foram calculadas 10 métricas, com base em dados recolhidos na bibliografia, para 72 espécies de 16 sistemas estuarinos (costas O e S de Portugal). A validação do índice foi feita por meios externos. Este é o primeiro índice desenvolvido para avaliar a importância conservacionista de peixes estuarinos e permite identificar as espécies com maior e menor prioridade de conservação, auxiliando o planeamento da conservação de peixes estuarinos

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Chapter 1

General Introduction

“Before I speak, I have something important to say. “

Groucho Marx (1890-1977)

General Introduction

Estuarine systems are considered to be some of the worlds' most productive and valuable aquatic ecosystems (Costanza et al. 1997). Marine species attain here a greater growth potential, confirming thus the importance of these systems as nursery areas for comercial importante species and to other taxa (Marchand 1980; Costa & Bruxelas 1989; Blaber et al. 2000; Gillanders et al. 2003; Le Pape et al. 2003; Able 2005). Nevertheless, this importance has not yet been properly assessed (Cabral & Costa 2001; Erzini et al. 2002; Martinho 2005). Estuaries can also act as a nutrient and organic matter provider to the neighbouring coastal areas, increasing the primary and secondary production. As transitory systems, they are recognised as part of the continental coasts having extreme biological and economic importance and being stage to a large number of human activities (Houde & Rutherford 1993; Cooper et al. 1994; Marques et al. 2004).

Estuaries are very complex and are ruled by environmental gradients, most of them derive from the fact that marine salt water and river freshwater meet here (Haedrich 1983). Their mixture creates a wide range of gradients that favours the recruitment of a vast array of species with different physical and trophic characteristics/needs (Sánchez & Raz-Guzman 1997; Harris et al. 2001; Kimmerer et al. 2001). The greater content in organic matter and nutrients present in these systems favours the establishment of benthic trophic webs, resulting in a food availability increase for the juveniles of several species (Peterson et al. 2000; Grall & Chauvaud 2002; Salen-Picard et al. 2002). In fact, fish juveniles have a particular tendency to use estuaries, largely due to the advantages provided by these systems in terms of growth, survival, food availability and protection from predators (Haedrich 1983; Miller et al. 1985; Lenanton & Potter 1987).

The fish fauna present at the estuaries experience a temporal and a spatial variation. These variations occur mostly due to species migrations (Elliott et al. 1990; Maes et al. 1998), to fish zonal distribution in relation to their saline tolerance (Loneragan et al. 1986; Henderson 1989) and to species preferences regarding sediment, vegetation and food (Elliott & Dewailly 1995; Marshall & Elliott 1998). The predatory pressure also influences habitat choice (Laegdsgaard & Johnson 2001). McHugh (1967) proposed a categorization of the species present in the estuary based on their usage of the system: freshwater fishes that occasionally visit the estuary, truly estuarine species,

diadromous species, marine species that regularly visit the estuary as adults, marine species that use the estuary as a nursery area and marine occasional species.

Public and political consciousness towards conservation and biodiversity is increasing, demanding more impact assessment studies, ecological monitoring programs and management plans. Conservation biology tries to keep the biodiversity level by preserving the endangered species that otherwise could be lost. Biodiversity in relation to ecosystem function is one of the emerging areas of research in environmental biology and very little is known about it at national or international level (Singh & Sharma 1998). Protection of locations with high species richness is an effective way to protect overall biodiversity and sustain key ecological functions (Scott et al. 1987; Myers et al. 2000). In other words, species richness is assumed to be an indicator of conservation value (e.g., Meir et al. 2004).

In an effort to conserve biodiversity, resources are often directed towards protecting rare species in the belief that these are the species most at risk of extinction (Gaston 1994). In the absence of detailed population viability analyses, some of the main indicators used for the assessment of extinction risk are rarity, rates of decline and degree of population fragmentation (Hartley & Kunin 2003). Qualitative methods for categorizing species according to risk (e.g., Fitter & Fitter 1987) are appealing because they are simpler and less reliant on difficult-to-gather data. However, such methods suffer from a nearly complete reliance on expert opinion and from the difficulty in assigning species to distinct categories of risk (Mace & Lande 1991). Adoption of quantitative methods has been hampered by lack of data (Ceballos & Navarro 1991; Mace 1994) but their objectivity and repeatability are desirable attributes not found in qualitative methods (Todd & Burgman 1998).

The need for a priority-setting process is driven by limited conservation resources that can only be attributed to a small percentage of all possible species in any given geographic area. This need has led to the development of practical systems for categorizing and assessing the degree of vulnerability of various components of biodiversity, particularly vertebrates (e.g., TNC 1988; Millsap et al. 1990; Master 1991; Martin 1994; Stotz et al. 1996). Prioritization systems in use or proposed vary greatly in what factors are concerned, how these factors are scored, weighted and integrated, and how the resulting information is presented and used (Mehlman et al. 2004)

Species conservation demands an intrinsic and vast knowledge of the species biology and ecology, which is impossible for most of the species. To overcome this

great information demand there are multimetric indices that can prioritize species in accordance to their perceived extinction risk (e.g., Millsap et al. 1990). These methodologies are simple to use, they rank the species by conservation priority based on some easily measured metrics.

Regarding the estuarine fish species, there is not an index that allows species ordination by conservation priority. Thus the creation of such an index would be an innovation. A conservation priority index would allow decision makers to understand the data and to take a correct decision, more in line with the real state of the systems and of the species. The aim of this work is to propose a multimetric index for the conservation of the Portuguese estuarine fish species, using the available bibliographic information on the species, resulting in a complement to the red list of endangered species (Cabral et al. 2005), which does not consider most of these species.

Chapter 2

Conservation Priority Index for Estuarine Fish
(COPIEF).

(submitted)

Conservation Priority Index for Estuarine Fish (COPIEF).

Paulo Branco^{1,2}, José Lino Costa¹ and Pedro Raposo de Almeida^{1,3,§}

1 – Institute of Oceanography, Faculty of Sciences, University of Lisbon, Campo Grande, 1749-016 Lisbon, Portugal

2 – Department of Forestry, Institute of Agronomy, Technical University of Lisbon, Tapada da Ajuda, 1340-017 Lisbon, Portugal

3 – Department of Biology, University of Évora, Apartado 94, 7002-554 Évora, Portugal.

§ - Corresponding author – E-mail: pmraposo@mail.telepac.pt; Tel. +351266760800; Fax. 351266760914

Abstract:

Public awareness regarding environmental issues has been increasing in the last decades. The increasing number of impact assessment studies, management and conservation plans, as well as ecological monitoring studies, demand new and more efficient techniques. Indices are an important tool aiding biologists in studies and allowing an easier comprehension of the data by managers, decision makers and general public. This study presents the first multimetric index for establishing a hierarchical ordination of the conservation priority of estuarine fish species using, for that, 72 species from 16 estuarine systems (W and S coasts of Portugal). The index is composed by ten metrics, comprising species' life traits, distribution and population trends. The information needed to score each metric was gathered from bibliography and the index validation was done by external means. This methodology allowed the definition of the fish species more in need of conservation planning, and those less prone to extinction in estuarine systems. The proposed index will fill a gap

in the knowledge and contributes with a useful tool to the scientific community and to the decision makers, being a breakthrough on estuarine fish species conservation planning.

Keywords: Conservation of biodiversity, multimetric index, fish species, biodiversity, estuaries and coastal lagoons, Portugal

Introduction

Coastal transition systems, such as estuaries and coastal lagoons, are vital ecosystems to marine fish stocks and a passageway for diadromous migrants. Despite their importance, estuaries are also amongst the most modified and threatened aquatic ecosystems in the world (Blaber et al. 2000), suffering from severe human impacts that imperil their natural ecological function, limiting thus the continuity of the systems' health and viability (Goldberg 1995; Costa et al. 2002; Kennish 2002).

The conservation of the natural systems is one of the most important concerns of today's world. So, conservation planning must be undertaken with extreme rigor (Fleishman et al. 2006). The maximum goal of conservation biology is to maintain the biodiversity level (Linguist 2007), once it has an incredible value to human welfare. "Endangered species are elements that contribute to a region's biological diversity and whose regional extinction represents a measurable and potentially irretrievable loss" (Maiorano et al. 2006).

Species richness has always been used as an indicator of ecological condition or conservation value, mostly due to its information availability over long periods of time and on distinct locations (Fleishman et al. 2006). The use of species richness may be more related to the information availability and to the biologists desire to produce lists than to the real ecological significance (Fleishman et al. 2006). The use of simple lists has proved to be unsuccessful, conservation wise, once it considers each species as an equivalent unit, disregarding its individual ecological role (Piraino et al. 2002). In fact, the establishment of conservation priorities requires a vast range of ecological/biological values (Fleishman et al. 2006).

Many endangered species are so because of the loss of critical habitat to one or more life stages (e.g., Wilcove et al. 1998). The ecosystems natural populations'

maintenance is an effective way to maintain the biodiversity (Balmford et al. 1996; Redford & Richter 1999; Groves 2003; Rosenzweig 2003) as it is the habitat protection.

With the continuous increase in public and political awareness towards conservation (Fletcher 2005) and the profusion of impact assessment and management studies, the development of tools and techniques which help the professionals to perform a more accurate, economic and less time consuming work is very important. Conservation biology is an applied science and, as such, is based on values (Meffe & Carrol 1994; Barry & Oelschlaeger 1996). The use of indexes in this kind of works is increasing (Spellerberg 1993; Olivier & Beattie 1994), since they simplify the data (Graça & Coimbra 1998; Hartwell 1998; Harris & Silveira 1999; Ladson et al. 1999) and simplify the decision making, maintaining the scientific accuracy (Paul 2003). This simplification (Ranasinghe et al. 2002) allows a better understanding of the problems by the decision makers (Knapp et al. 2003), enabling them to take correct decisions in accordance with the real state of the communities (Karr & Chu 1999). As for any model, the merit of an index is not only measured by the reality representation fidelity, but also in relation to its interpretation capability (Gómez Orea 2002).

Quantitative methods for prioritising species regarding their conservation needs, range from easily measured simple variables (e.g., Freitag & Van Jaarsveld 1997; Cofré & Marquet 1999) to more sophisticated methods (e.g., Millsap et al. 1990). This kind of methods has had a limited expression on conservation due to the lack of data (Ceballos & Navarro 1991; Mace 1994) or due to the difficult-to-gather data, but the capability of being repeated and their intrinsic objectivity are desirable qualities not found in the traditional qualitative methods (Todd & Burgman 1998). Moreover a quantitative method can be compared across countries or across distinct geographic areas (Ceballos & Navarro 1991). A method which offers an ordination of the species, conservation wise, aids in the distribution of limited conservation efforts/resources (Mace 1995; Reyers et al. 1998; Mehlman et al. 2004), maximizing thus the beneficial outcomes. Prioritising species based on their perceived endanger and conservation needs is a popular methodology in conservation biology (Rabinowitz 1981; TNC 1988; Mace & Collar 1994; Carter et al. 2000; Stein et al. 2000).

The information gaps and lack of knowledge of some species or of some life traits characteristics is a threat to the biodiversity and imperils conservation, once there are not valid values to base the conservation efforts on (Mares 1986). Even the

identification of these knowledge gaps of the species biology is an important step towards species conservation (Cofré & Marquet 1999).

There is not, as far as our knowledge, an index which allows a fish species ordination by conservation priority. Thus, the creation of such an index would be an innovation, being the information gathered from the index a complement to the Portuguese Red List of Endangered Species (Cabral et al. 2005), which has considerable knowledge gaps concerning these species, and would create an hierarchical ranking instead of a simple list. A conservation priority setting system will be as reliable as the amount of knowledge available on the species contemplated. Nonetheless, a system with a great baseline information demand tends to have a smaller application spectre (Palmeirim et al. 1994; Harcourt & Parks 2003). A multimetric index with the capacity of ranking species in regard to their conservation priority has to have metrics which integrate pertinent information to the species ordination (Jackson et al. 2000), and has to comprise the biological sensitivity of the species as well as its population relevance (Palmeirim et al. 1994). This work purposes the enunciation of the first multimetric index for the conservation priority of estuarine fish species.

Materials and methods

Estuarine systems

In this work we studied 16 Portuguese systems (Fig. 1). These are the most important estuaries and open and semi-opened coastal lagoons in both the West and South (Algarve) coasts of Portugal. The spatial coverage allowed by the use of these systems gives a good spatial confidence to the index. Portugal is a biogeographic transitional zone. The Cape Carvoeiro constitutes the border between two zoogeographic zones, the Lusitanian and the West African, these zoogeographic zones represent also two distinct coastal climatic divisions, the Western Temperate and the Subtropical (Hayden et al. 1984). The estuarine systems present different characteristics between the two geographic zones so the systems were separated into northern and southern systems, being the Tagus estuary the first southern estuary.

Fish species

In the formulation of this multimetric index we considered 72 fish species. These species were selected based on a bibliographic collection of the fish fauna composition of the 16 estuarine systems, independently. To assure that just the estuarine species (the species that use the estuary frequently and not occasionally) were assessed, only the species that were present in more than 50% of the northern or the southern systems were included in the analysis. Nevertheless, species with especial importance/relevance in one or more systems were included even if they did not fulfil the above stated premises. The species that are now considered regionally extinct in Portugal (Cabral et al. 2005) were not included in the analyses.

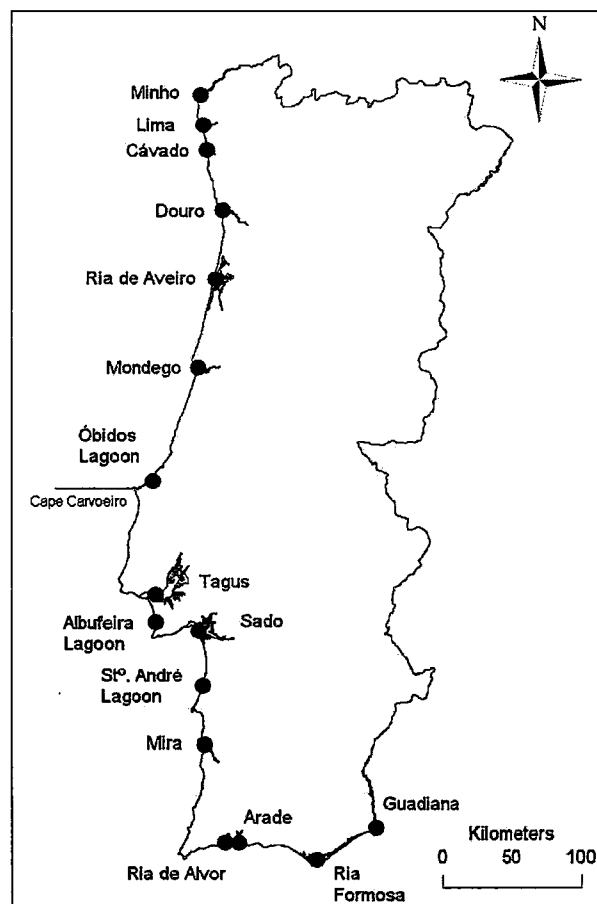


Figure 1 – Location of the 16 studied estuarine systems and of the Cape Carvoeiro that separates the northern from the southern estuarine systems.

Information gathering

The information regarding the potential metrics was gathered, selected and derived from a detailed and extensive bibliographic review (annex I). There was also punctual information facilitated by specialists of certain species or species groups.

Metric selection

In this numeric approach to the conservation of estuarine fish, several biological metrics that allow an ordination of the species in accordance to their conservation priority, were enunciated. These metrics facilitate a comprehension of the species relative sensibility. The list of potential metrics derived both from the bibliography as from the authors' personal knowledge of estuarine species. After the bibliographic search for the information on biological metrics for each species some of the original metrics were excluded from the index. The first criterion for metrics exclusion was the lack of information for many species. The second criterion for exclusion was the information redundancy input of the different metrics, which was evaluated using the Spearman's rank correlation test (Sokal & Rohlf 1995), that lead to the exclusion of one metric when it was highly correlated ($r > 0.8$) with another, since both metrics were contributing with the same information input to the overall species hierarchy. The final 10 metrics selected represent the life trait of the species, their distribution and population trends.

Metric scoring

The scoring of each metric was reached taking in consideration the value attained by all the species in that metric, to better fit the score classes within the metrics' value range, and attributing a higher score to the class with greater conservation priority. The scoring process had some simple premises, the metrics had to be scored in relation to the adult fish and its survival probability. The metrics scores varied from zero (low conservation priority) to ten (high conservation priority). As the metric scoring values correspond to a class interval, when a species had an information gap it was filled, if possible, with that metrics information of a relative species (i.e., a species of the same genus or family) resulting thus in a surrogating value. When there was an impossibility to replace knowledge gaps, the value five was attributed to that

metric. The value “five” was chosen as it is the central value in the metric score spectrum and it does not tend to pull the overall index score down or upwards.

Metrics

Fenology

This metric (Table 1) is based on the estuarine temporal occupation pattern of the species in Portugal. This usage evaluation was done both with resource to bibliography and with the authors and species specialists' knowledge of the species and their behaviour in Portuguese estuarine systems. The higher conservation values were attributed to the species which use the estuarine area for long periods of time, i.e., the resident species, being the lowest priority assigned to the marine species which only visit the estuary briefly for protection or foraging. The premise underlined in this scoring table is that a resident species is subjected to the same and more survival hazards as a species using the estuary only during a part of the life cycle (nursery or diadromous species) or just as an increment to their usual habitat (marine species).

In this metric the term nursery is applied to species which make an extensive use of estuarine systems as growth areas for juveniles (but that do not complete their life cycle in the estuary), since these systems grant some degree of protection from the predators and are resource richer environments.

Distribution

Regional distributional range (Table 1) – The West coast of Portugal is a zoogeographic transition region (Hayden et al. 1994) and, as such, many fish species have in Portugal their latitudinal distribution limit. Bearing this in mind, a species which finds in Portugal its distributional limit has to have a higher conservation value than one that does not. A species that finds here its southern limit has to be more protected than one that finds here its northern limit of distribution, due to the climatic changes (rising temperatures) and the noticeable latitudinal changes of the species distributions in the last decades. This scoring methodology is based on the premise that a species with its distribution limit in Portugal is more likely to disappear from this coast than a species without its distribution limit in Portugal.

Local distributional range (Table 1) – This metric refers to the dispersion of a species in Portuguese estuarine systems. The fundamental premise here is that a distribution restricted species has to have a higher conservation priority than a national widespread one. By the same principal referred to the regional distribution, species with a clear northern distribution were ranked higher than southern species.

Intra-estuarine zonation (Table 1) - This metric refers to the space usage done by the species in the estuarine system. A species that uses the entire estuary has lower conservation priority than one that uses only a part of the estuary, or a specific habitat. The species which occur at the mouth of the estuary have a greater capability to use the marine environment, having thus greater habitat availability, not deserving a high conservation value. The species with a seagrass beds zonal distribution had the highest score in this metric since this habitat availability is narrower.

Trophic ecology

This metric (Table 1) is based on the species trophic level and has as a premise the food availability in the system, and the energy needed to harvest that food item. Bearing this in mind, the species which consume less available items and feed at a higher trophic level have a greater score than the species whose food items are abundant. That is why the species with some trophic plasticity have a lower score than more specialized ones. Piscivores and parasites have the highest score value followed by species which are mainly piscivores but which have an important invertivore component in their diet. Following these species in the metric score, are the mainly invertivore species, then the species which feed primarily on invertebrates but vegetated matter is an important food item for them. The lower two metric classes are composed by herbivores and by the plankton and detritic feeders.

Maximum size

The metric (Table 1) scoring is, in this case, based on the size of the animal; large animals score higher than small ones. The underlined premise is that larger animals need more resources as well as greater vital areas. This is a valid simple assessing way to evaluate the vital area extension, once that this information is not available for many species. It was thus considered that species which attained larger

sizes are more vulnerable than species with smaller sizes. For every species it was used the highest valid size found in the bibliography.

Reproduction strategy

Age at first maturity (Table 1) - The ranking scores in this metric relate to the probability of a successful reproduction. The premise is that a species with a younger age at first maturity has a higher probability of survival and of reproductive success than a species which matures later in the life span, scoring thus a lower conservation priority. For every species it was used the highest valid age at first maturity found in the bibliography.

Reproductive guilds (Table 1) - The reproductive guild metric is directly related to the degree of protection done by the adults to the hatch, implying thus greater or smaller energy consumption by the progenitors. The premise here is that a higher degree of protection leads to the increase in energy expenditure; therefore, there is a higher vulnerability of these species, scoring a higher conservation value.

Spawning frequency (Table 1) - The spawning frequency relates to conservation in the way that a species which spawns several times in its life is more likely to overcome, as a species, adverse conditions for the survival of the juveniles/eggs than a species which only spawns once in a lifetime, fitting, thus, in a lower scoring class.

Reproductive guilds and spawning frequency were pooled into one metric to avoid overweighting reproduction strategy.

Species national trends

This was a very hard metric (Table 1) to calculate, once this information is largely unknown by the scientific community. So, there were a lot of information gaps and these were filled using information from the National Fisheries Department (DGPA) and abundance data found in the bibliography. These data allowed the determination with some degree of certainty the probable trends of the species in the Portuguese territory. Nevertheless, the scoring classes for the extrapolated trends were

kept nearer the “unknown” class (Table 1) than of the extremes of the metric, granting thus a minor interference of the extrapolation of the data on the final index result. Even though the difficulty in determining the species trends, this metric was kept in the index, once it is the only one that allows an integration of the actual state of the population.

Taxonomic importance

This metric (Table 1) is very precise and direct, it uses only the systematic position of a given species. The premise underlined here is that a species which is the sole representative of its family or genus is of greater conservation value than a species which shares its higher taxonomic level with more species. The loss of a family or genus is of greater importance to the biodiversity than the loss of a species.

Index calculation

The information provided by the metrics used to enunciate the index only makes sense when pooled together. Only the metrics’ values integrated grant a unified view on the species relative sensibility. The Conservation Priority Index for Estuarine Fish (COPIEF) is a holistic way to define conservation priorities. The index is calculated by averaging the values attained by a species in all the metrics.

Validation

To assess the robustness and accuracy of the index, it was validated by external means. The first validation step was to validate our results with the criteria of the IUCN (IUCN 2001). The second route of validation was supported by experts’ knowledge. In this second validation methodology the species list was delivered to five specialists of Portuguese estuarine fish (see acknowledgments). The specialists were asked to give a priority conservation value from 1 to 3 (being 1 the lowest priority and 3 the higher) in accordance with their knowledge of the species. The specialists’ rankings were averaged to get an overall accordance specialists’ index. Then this ranking was compared (Spearman’s rank correlation test) to the one accomplished by the purposed index to assert if there were common views between the index and the experts informed insight.

In order to further validate the index results, this multimetric methodology was applied on a recently regionally extinct species (*Acipenser sturio* Linnaeus, 1758), just

to make an index estimative. Theoretically, an extinct species should attain a high conservation priority score on the proposed index.

Table 1. Metrics ranking scores for the Conservation Priority Index for Estuarine Fish (COPIEF)

Metric	Score
Fenology	
Resident	10
Nursery + Diadromous	5
Marine	0
Regional Distribution	
Southern distribution limit	10
Northern distribution limit	5
Widespread	0
Local Distribution	
Restricted	10
Predominantly Northern	7
Predominantly Southern	4
Widespread	0
Intra-estuarine Zonation	
Mainly at seagrass beds	10
Mainly at middle or upper estuary	5
Mainly at the estuary mouth or all the estuary	0
Trophic Ecology	
Parasite + Piscivore	10
Piscivore and Invertivore	8
Invertivore	6
Invertivore and Herbivore	4
Herbivore	2
Plankton feeders + Detritic feeders	0
Maximum Size	
> 100 cm	10
71 – 100 cm	8
51 – 70 cm	6
21 – 50 cm	4
11 – 20 cm	2
5 – 10 cm	0
Age at First Maturation	
> 10 years	10
9 – 10 years	8
6 – 8 years	6
3 – 5 years	4
1 – 2 years	2
< 1 year	0
Reproduction Strategy	
Reproductive Guilds	
Internal transportation	5
Parental protection	4
Benthic eggs (sticky/shells)/nesting	2
Pelagic spawner	0

(Continued on next page)

Table 1 (continued)

Metric	Score
Spawning Frequency	
Smelparous	5
Iteroparous	0
National Population Trends	
Decreasing	10
Unknown but possibly decreasing	7
Unknown	5
Unknown but possibly stable or increasing	3
Increasing or stable	0
Taxonomic Relevance	
Sole family representative in Portuguese estuaries	10
Sole genus representative in Portuguese estuaries	5
Multispecific genus in Portuguese estuaries	0

Results

Although the difficulty to find data to correctly score the species in all the metrics (Table 2), only five values (0.63%) out of 792 were scored as unknown, due to lack of information on the species or their relatives. The Spearman rank correlation test showed that there were not any high correlations between the chosen metrics, meaning that there is not information redundancy within those metrics.

The index score varies from 0.9 to 7.1, having a wide score spectrum and allowing a good species segregation. The upper and lower ends of the index distribution are clearly isolated from the middle bulk of the ranking, which groups all the species that present only slight score variation, resulting in a continuous decrease instead of a step fall.

Validation

This multimetric index ranking system had to be validated by external means. It was tried to validate the index with the IUCN criteria but the kind of information needed to do that does not exist. So, in view of this, it was impossible to pursue this validation route. The way found to accomplish this task was to request the assistance of five Portuguese estuarine fish species experts. The index ranking and the experts ranking are positively correlated with high significance ($r = 0.49$, $N = 72$, $p < 0.001$).

To corroborate this finding it was calculated the conservation priority index value for the regionally extinct species *A. sturio* which scored 7.4. This value would place it in the first ranking position just before the salmon (*Salmo salar* Linnaeus, 1758).

Table 2. Species ranked by the value attained in the Conservation Priority Index for Estuarine Fish (COPIEF), with indication of the score attained by each species in each metric.

#	Species	FEN	RD	LD	ID	TE	MS	FM	RG	RF	NT	TAX	COPIEF
1	<i>Salmo salar</i> Linnaeus, 1758	5	10	10	5	8	10	6	2	5	10	0	7.1
2	<i>Anguilla anguilla</i> (Linnaeus, 1758)	5	0	0	5	8	10	10	0	5	10	10	6.3
3	<i>Lampetra fluviatilis</i> (Linnaeus, 1758)	5	0	10	0	10	4	6	2	5	10	5	5.7
3	<i>Petromyzon marinus</i> Linnaeus, 1758	5	0	7	0	10	10	6	2	5	7	5	5.7
5	<i>Argyrosomus regius</i> (Asso, 1801)	5	0	10	0	10	10	4	0	0	7	10	5.6
6	<i>Conger conger</i> ([Artedi, 1738] Linnaeus, 1758)	5	0	0	0	8	10	10	0	5	7	10	5.5
6	<i>Raja undulata</i> Lacepède, 1802	5	0	4	0	6	8	8	4	0	10	10	5.5
8	<i>Salmo trutta</i> Linnaeus, 1758	5	0	10	5	8	10	4	2	0	10	0	5.4
9	<i>Gasterosteus aculeatus</i> (Linnaeus, 1758)	5	0	7	5	6	2	2	2	5	7	10	5.1
10	<i>Alosa alosa</i> (Linnaeus, 1758)	5	0	0	5	8	8	8	0	5	10	0	4.9
10	<i>Ciliata mustela</i> (Linnaeus, 1758)	5	10	7	0	6	4	2	0	0	5	10	4.9
12	<i>Dicentrarchus labrax</i> (Linnaeus, 1758)	5	0	0	5	8	10	8	0	0	10	0	4.6
13	<i>Plathichthys flesus</i> (Linnaeus, 1758)	5	0	0	5	8	6	4	0	0	7	10	4.5
14	<i>Alosa fallax</i> (Lacepède, 1803)	5	0	0	5	8	6	8	2	0	10	0	4.4
14	<i>Halobatrachus didactylus</i> (Bloch & Schneider, 1801)	10	0	4	0	8	6	2	4	0	0	10	4.4
14	<i>Monochirus hispidus</i> Rafinesque, 1814	10	5	4	5	6	2	2	0	0	5	5	4.4
17	<i>Ammodytes tobianus</i> (Linnaeus, 1758)	10	10	0	0	0	2	2	2	0	7	10	4.3
17	<i>Serranus hepatus</i> (Linnaeus, 1758)	5	0	4	0	8	4	5	0	0	7	10	4.3
19	<i>Callionymus lyra</i> Linnaeus, 1758	5	0	0	0	6	4	6	0	5	5	10	4.1
19	<i>Symphodus bailloni</i> (Valenciennes, 1839)	10	0	0	10	6	4	2	4	0	5	0	4.1
21	<i>Trigla lucerna</i> Linnaeus, 1758	5	0	0	0	8	8	2	0	0	7	10	4.0
22	<i>Sarpa salpa</i> (Linnaeus, 1758)	5	0	0	10	2	6	4	0	0	7	5	3.9
22	<i>Symphodus cinereus</i> (Bonnaterre, 1788)	10	0	0	10	6	2	2	4	0	5	0	3.9
24	<i>Dicentrarchus punctatus</i> (Bloch, 1792)	5	0	4	0	8	6	8	0	0	7	0	3.8
24	<i>Diplodus puntazzo</i> (Cetti, 1777)	5	5	4	0	4	6	2	0	0	7	5	3.8
26	<i>Nerophis lumbiciformis</i> (Jenyns, 1835)	10	0	0	10	0	2	0	5	0	5	5	3.7
27	<i>Scorpaena notata</i> Rafinesque, 1810	0	0	4	0	8	4	5	0	0	5	10	3.6
28	<i>Belone belone</i> (Linnaeus, 1761)	0	0	0	0	8	8	2	2	0	5	10	3.5

(Continued on next page)

Table 2 (continued)

#	Species	FEN	RD	LD	ID	TE	MS	FM	RG	RF	NT	TAX	COIEF
28	<i>Scophthalmus rhombus</i> (Linnaeus, 1758)	0	0	0	0	8	8	2	0	0	7	10	3.5
28	<i>Trachurus trachurus</i> (Linnaeus, 1758)	0	0	0	0	8	6	4	0	0	7	10	3.5
31	<i>Hippocampus hippocampus</i> (Linnaeus, 1758)	10	0	0	10	0	2	0	5	0	7	0	3.4
31	<i>Hippocampus ramulosus</i> Leach, 1814	10	0	0	10	0	2	0	5	0	7	0	3.4
31	<i>Syngnathus acus</i> Linnaeus, 1758	10	0	0	10	0	4	0	5	0	5	0	3.4
31	<i>Syngnathus typhle</i> Linnaeus, 1758	10	0	0	10	0	4	0	5	0	5	0	3.4
35	<i>Solea senegalensis</i> Kaup, 1858	5	0	0	5	6	6	4	0	0	7	0	3.3
35	<i>Solea solea</i> (Linnaeus, 1758)	5	0	0	5	6	6	4	0	0	7	0	3.3
37	<i>Buglossidium luteum</i> (Risso, 1810)	0	0	10	0	6	2	4	0	0	5	5	3.2
37	<i>Pagellus bogaraveo</i> (Brünnich, 1768)	0	0	0	0	8	6	6	0	0	7	5	3.2
37	<i>Pagrus pagrus</i> (Linnaeus, 1758)	0	0	0	0	8	8	4	0	0	7	5	3.2
40	<i>Atherina boyeri</i> Risso, 1810	10	5	0	5	0	2	2	2	0	5	0	3.1
40	<i>Gobius paganellus</i> Linnaeus, 1758	10	0	0	0	6	2	4	4	0	5	0	3.1
40	<i>Microchirus azevia</i> (Capello, 1867)	0	5	4	0	6	4	4	0	0	3	5	3.1
40	<i>Mugil cephalus</i> Linnaeus, 1758	5	0	0	0	0	10	6	0	0	5	5	3.1
44	<i>Boops boops</i> (Linnaeus, 1758)	5	0	0	10	2	4	4	0	0	0	5	3.0
44	<i>Mullus surmuletus</i> Linnaeus, 1758	5	0	0	0	6	4	2	0	0	3	10	3.0
44	<i>Parablennius pilicornis</i> (Cuvier, 1829)	10	5	0	0	4	2	2	2	0	5	0	3.0
44	<i>Pomatoschistus marmoratus</i> (Risso, 1810)	10	5	0	0	6	0	2	2	0	5	0	3.0
44	<i>Pomatoschistus microps</i> (Krøyer, 1838)	10	0	0	0	6	0	2	4	5	3	0	3.0
49	<i>Aphia minuta</i> (Risso, 1810)	10	0	0	0	0	0	2	2	5	5	5	2.9
49	<i>Engraulis encrasicolus</i> (Linnaeus, 1758)	10	0	0	0	0	2	2	0	0	5	10	2.9
49	<i>Parablennius gattorugine</i> (Linnaeus, 1758)	10	0	0	0	4	4	2	4	0	5	0	2.9
49	<i>Syngnathus abaster</i> Risso, 1826	10	0	0	5	0	4	0	5	0	5	0	2.9
53	<i>SpondylIOSoma cantharus</i> (Linnaeus, 1758)	5	0	0	0	6	6	4	2	0	0	5	2.8
54	<i>Gobius Niger</i> Linnaeus, 1758	10	0	0	0	6	2	2	2	0	5	0	2.7
54	<i>Sparus aurata</i> Linnaeus, 1758	5	0	0	0	4	6	4	0	0	3	5	2.7
56	<i>Diplodus bellottii</i> (Steindachner, 1882)	5	5	4	0	6	4	2	0	0	0	0	2.6
57	<i>Chelon labrosus</i> (Risso, 1826)	5	0	0	0	0	6	4	0	0	5	5	2.5
57	<i>Diplodus cervinus</i> (Lowe, 1841)	0	0	4	0	4	6	4	0	0	7	0	2.5

(Continued on next page)

Table 2 (continued)

#	Species	FEN	RD	LD	ID	TE	MS	FM	RG	RF	NT	TAX	CO	PIEF
57	<i>Pomatoschistus minutus</i> (Pallas, 1770)	10	0	0	0	6	2	0	4	0	3	0	2.5	
57	<i>Pomatoschistus pictus</i> (Malm, 1865)	10	0	0	0	6	0	2	2	0	5	0	2.5	
61	<i>Liza saliens</i> (Risso, 1810)	5	0	0	5	0	4	4	0	0	5	0	2.3	
62	<i>Diplodus sargus</i> (Linnaeus, 1758)	5	0	0	0	4	4	2	0	0	7	0	2.2	
63	<i>Arnoglossus laterna</i> (Walbaum, 1792)	0	0	0	0	8	4	4	0	0	5	0	2.1	
63	<i>Arnoglossus thori</i> Kyle, 1913	0	0	4	0	8	2	2	0	0	5	0	2.1	
63	<i>Echichthys vipera</i> (Cuvier, 1829)	0	0	0	0	0	2	4	0	0	5	10	2.1	
63	<i>Solea lascaris</i> (Risso, 1810)	0	0	0	0	6	4	4	0	0	7	0	2.1	
67	<i>Liza aurata</i> (Risso, 1810)	5	0	0	0	0	6	4	0	0	5	0	2.0	
67	<i>Liza ramado</i> (Risso, 1826)	5	0	0	0	0	6	4	0	0	5	0	2.0	
69	<i>Diplodus annularis</i> (Linnaeus, 1758)	0	0	0	0	6	4	2	0	0	7	0	1.9	
70	<i>Diplodus vulgaris</i> (E. Geoffroy Saint-Hilaire, 1817)	5	0	0	0	6	4	2	0	0	0	0	1.7	
71	<i>Sardina pilchardus</i> (Walbaum, 1792)	0	0	0	0	0	4	4	0	0	0	5	1.3	
72	<i>Atherina presbyter</i> Cuvier, 1820	0	0	0	0	0	2	0	2	0	5	0	0.9	

Acipenser sturio Linnaeus, 1758 (*)

5 10 10 5 6 10 6 2 0 10 10 7.4

FEN – Fenology; RD – Regional Distribution; LD – Local Distribution; ID – Inter-estuarine Distribution; TE – Trophic Ecology; MS – Maximum Size; FM – Age at First Maturation; RG – Reproductive Guilds; SF – Spawning Frequency; NT – National Population Trends; TAX – Taxonomic Relevance; (*) – Regionally extinct.

Discussion

The proposed methodology allowed the ranking of the Portuguese estuarine fish species. This ordination enabled the identification of the next species up for extinction, as it did for the species with least conservation concern. The top 11 ranked species are species with real conservation hazards that, due to their life trait characteristics, their distribution and their population trends, are on the road to a possible regional extinction. These 11 species are ranked higher in the conservation hierarchy because they feed at a high taxonomic level, have a scattered or restricted local distribution and most of them have a decreasing national population trend, attain a high body size and are semelparous. All the species ranked in the top 11 positions are diadromous or use the estuarine system as a nursery ground, having all of them the same fenology score. From these 11 species, seven are listed as endangered species in the Portuguese Red List of Endangered Species (Cabral et al. 2005), the four remaining are marine species which are not contemplated in the former list, this reveals that COPIEF is a valid methodology. These 11 species are the ones in which the conservation plans should be focused. The bottom ranked species are species which have a greater capacity to persist as a species, not needing an urgent conservation plan dedicated to them. On the other hand, it is possible to detect in the lower end of the ranking a step in the index values which allows a separation of the 12 last species. These species fit in such a low ranking position because they have a widespread regional, local, and intra-estuarine distribution, and they spawn several times during their life, having some other favourable life traits. The combination of favourable characteristics places these species in a low concern position, once they grant them a certain resistance against extinction pressures. Nevertheless, this does not mean that they are endless resources to be exploited. They can in fact be exploited with a certain degree of safety but, due to global changes and environmental pressures their hierarchical position may be altered in a way that they can not be recovered from. So, even species with a low conservation priority, still have to be integrated in general conservation strategy and, when exploited it has to be done with caution and in a sustained way. The low ranking species must never be neglected since they maintain fisheries and are on the base of the food web, contributing for the existence of various estuarine and coastal processes (Piraino et al. 2002). For instance *Sardina pilchardus* (Walbaum, 1792) represented, in the last 10 years, $83.6 \pm 12.1\%$ of all the fish captured by the Portuguese fishing fleet (DGPA 1997, 1998, 1999, 2000,

2001, 2002, 2003, 2004, 2005, 2006) and *Liza ramado* (Risso, 1826) is one of the most important species in what regards to estuarine biomass availability (Almeida et al. 1993) and have a valuable ecological role in these systems. So, even though their low ranking, these species are very important for the maintenance of the coastal biodiversity and stocks, being essential links of the systems' processes.

An index such as COPIPEF has to pass through a validation process to allow a certain confidence in the results. The fact that the insight of five independent experts, without any knowledge of the methodology followed to enunciate this index, is positively correlated with the ranking positions of the species attained by the use of the proposed index, is a clear indication of the index closeness to the true conservation priorities. The correlation value is quite high if considered that the specialists based their opinion on the known or perceived population trend of each species; and that each specialist is more knowledgeable of certain species or certain species groups. This results in an external validation of the index, confirming thus the accurateness of this multimetric approach to the conservation priority of the estuarine fish species. The index value for the *A. sturio* was calculated, and if it was not extinct in Portugal, it would be ranked as the top species most in need of a conservation strategy. This further reinforces the index accuracy, since that species did in fact go extinct. Bering this in mind, the top priority species really need to be protected, as should their habitats. The salmon is the next inline for the regional extinction, it is very fragile and its capacity to overcome future problems is limited. It has a high commercial value linked to a decreasing population trend and the southern limit of its distribution at the North of Portugal, being thus locally affected by the global climate changes. All these factors associated with unfavourable life trait characteristics, gravely imperil this species.

The impossibility of asserting the conservation status of the considered species by the IUCN criteria, due to the lack of information, reinforces the need of an index such as the one proposed in this paper. This index uses available information based on the species life traits, ecology, distribution and population trends, instead of population numbers, expected percentages of population decrease and distribution range limitation. Even the identification of knowledge gaps is the first step to conservation (Cofré & Marquet 1999). Identifying these gaps allowed a greater index validation as it will allow a classification, conservation wise, of certain species that otherwise would not be possible to classify. The Conservation Priority Index for Estuarine Fish gives new insight into estuarine fish species conservation, and it could function as a tool for fish

species conservation planning alongside with the Portuguese Red List of Endangered Species (Cabral et al. 2005). Vertebrate ecology is still incompletely known, so all ranking systems are imperfect (Báldi et al. 2001). Nevertheless, a ranking system such as the one developed in this work is a very important tool for ecologists. The work is easier and more doable than without this kind of methodologies. The proposed index is the first conservation priority ranking system for estuarine fish species and it is a breakthrough in terms of conservation biology. There are several indices of conservation priorities related to areas (Freitag et al. 1997; Bonn & Gaston 2005; Fox & Beckley 2005; Bani et al. 2006) or to other animal (Cofré & Marquet 1999; Polasky et al. 2001; Knapp et al. 2003; Mehlman et al. 2004; Keith et al. 2007) or plant groups (Martínez et al. 2006). Fish are under-represented in this kind of methodologies and there is the risk of neglecting fish conservation, or doing it inadequately. This index is thus of extreme importance, once that now there is the possibility to detect which species are most in need of a concentrated effort in conservation. A methodology such as this can work even on another level, it can be used to calculate sensitivity zones within estuaries and to identify relevant systems in terms of fish communities conservation. By adding the index values of the species found in each territorial area it is possible to have an local/system sensitivity index; the higher the value the more sensitive a zone is.

The Conservation Priority Index for Estuarine Fish establishes a new perspective in terms of fish conservation, being a new methodology applicable to other countries. “The analysis of priorities for conservation and threat of particular fauna is intrinsically a never-ending dynamic process. Species are continually moving among categories depending on the amelioration or increase in particular factors that impose on their numbers and distribution” (Cofré & Marquet 1999). Thus, further work must be done. The information gaps need to be filled to allow an even greater closeness of the index to the reality. And, as time goes by, the species ranking position, may alter themselves, due to climate changes, habitat degradation and reduction, and the alteration of anthropogenic pressures; so, periodical recalculation of the species index score must be done. The index should be widen to other areas creating a regional or a global index, it should also be applied, with slight modifications, to other environments such as coastal marine areas and freshwater systems

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Chapter 3

General conclusions

General conclusions

Conservation planning is one of the most urgent and difficult matters on the world's agenda. Assessing the species' risk of extinction is essential to regional conservation initiatives and to the allocation of limited resources for conservation (Master 1991; Mace 1995). The results of this study clearly show that it is possible to rank Portuguese estuarine fish species in accordance with their conservation priorities, maintaining a natural accurateness. In fact, the experts' knowledgeable insight is positively correlated with the list produced by the COPIPEF calculation, resulting in an external index validation. Furthering this validation route, it was calculated the index score for *A. sturio*, which is extinct, and it scored 7.4, placing this species at the top of the list. This fact acts as a second validation, once that the most endangered species did in fact go extinct. Of this validation methodology arises extra conservation concerns to *S. salar*, which scored 7.1. The unfavourable life traits characteristics of this species allied to a reduced local distribution, decreasing population trend and its great commercial value, gravely endanger the species survival, being this species on the edge of regional extinction. The ten other top ranked species are also extinction sensible, due to their unfavourable life traits, decreasing national trends, the fact that they feed at a high taxonomic level and that most of them attain a high body size and are semelparous. These eleven species urgently need to be actively preserved. This preservation can be made by designing specific conservation and restoration plans for the species and species' habitats. The bottom ranked species are more resistant to extinction pressures due to their widespread regional, local, and intra-estuarine distribution, and the fact that they spawn several times during their life, having additional favourable life traits. However, these species are keystones in the estuarine and coastal trophic webs and are responsible for maintaining the fisheries. Therefore, they must be considered in conservation plans elaboration. Only looking at the system as a whole can the protection of its' species and of its' processes be successful.

The methodology presented in this work provides an easy way to define conservation priorities and enables the ranking of estuarine zones and of estuarine systems (site-specific score) (Freitag et al. 1997) based on the fish species present, by adding their COPIPEF scores. The list of species and priorities produced by the index will allow to orientate and define the type of ecological research, as well as the contribution to perform local conservation activities (Martínez et al. 2005). This index,

being the first index for estuarine fish species, provides a starting point in terms of methodology and metrics for other country-based estuarine fish conservation priority indexes and it is a proof that multimetric approaches to estuarine fish are executable, viable and are a good alternative to local Red Lists specially as it responds to the need of conservation priority assessments (Dunn et al. 1999).

The next step following this study is the establishment of sensitivity zones on the 16 studied systems, simplifying thus, even further, the action of conservation planners and decision makers. The knowledge gaps that still remain after this information gathering should be filled with the continuous search of bibliography valid values. In order to provide a more accurate picture of the conservation priority of the estuarine fish species, the index should incorporate all new valid insights to the species knowledge, transforming the list presented here in an evolving organism, altering the ranking as the knowledge grows and the environmental and anthropogenic pressures change. In the future this index should be widened and transformed into a regional or global index and it should be applied, with some alterations, to the marine coastal area and to the freshwater environment.

Chapter 4

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Annex I

Species and reference number of the references used to score each species metrics

Alosa alosa (Linnaeus, 1758) – 23, 28, 55, 56, 71, 86, 90, 91, 97, 98, 99, 132, 178, 195, 197, 207, 224, 259, 272, 281, 302, 333

Alosa fallax (Lacepède, 1803) – 7, 13, 17, 24, 26, 29, 32, 34, 71, 91, 93, 96, 97, 98, 99, 101, 107, 108, 132, 160, 178, 207, 237, 259, 269, 333

Ammodytes tobianus (Linnaeus, 1758) – 32, 44, 91, 93, 96, 101, 134, 178, 195, 197, 204, 259, 266, 267, 268, 269, 278, 280, 318, 330

Anguilla anguilla (Linnaeus, 1758) – 12, 13, 17, 22, 24, 25, 26, 27, 32, 43, 54, 55, 62, 66, 71, 80, 91, 93, 95, 96, 98, 100, 101, 108, 110, 112, 132, 134, 136, 160, 178, 195, 197, 203, 224, 225, 237, 259, 269, 274, 278, 289, 303, 313, 315, 316, 317, 318, 330

Aphia minuta (Risso, 1810) – 13, 24, 32, 33, 77, 80, 93, 98, 101, 107, 108, 110, 174, 178, 190, 195, 197, 203, 213, 220, 234, 237, 259, 269, 294, 309, 313, 318

Argyrosomus regius (Asso, 1801) – 55, 68, 93, 95, 98, 205, 260, 313, 318

Arnoglossus laterna (Walbaum, 1792) – 13, 18, 62, 66, 69, 70, 93, 98, 108, 127, 157, 160, 203, 210, 234, 237, 242, 294, 313, 318

Arnoglossus thori Kyle, 1913 – 13, 18, 28, 62, 70, 80, 93, 101, 108, 152, 157, 224, 237, 242, 294, 313

Atherina boyeri Risso, 1810 – 41, 56, 65, 71, 98, 108, 110, 134, 160, 178, 195, 197, 234, 259, 262, 269, 307, 324

Atherina presbyter Cuvier, 1820 – 13, 17, 21, 24, 25, 26, 32, 38, 56, 88, 91, 93, 96, 98, 100, 101, 107, 108, 128, 132, 178, 195, 197, 214, 224, 225, 237, 250, 254, 259, 262, 269, 313, 318

Belone belone (Linnaeus, 1761) – 17, 32, 55, 62, 63, 66, 96, 98, 101, 108, 123, 153, 160, 173, 178, 195, 224, 225, 237, 259, 269, 278,

Boops boops (Linnaeus, 1758) – 12, 13, 24, 26, 44, 45, 66, 93, 98, 01, 107, 108, 110, 158, 159, 162, 195, 197, 204, 224, 225, 227, 237, 259, 294, 313

Buglossidium luteum (Risso, 1810) – 44, 66, 70, 96, 108, 110, 115, 160, 203, 244, 261, 280, 313

Callionymus lyra Linnaeus, 1758 – 32, 66, 82, 91, 93, 96, 98, 100, 101, 108, 110, 148, 178, 183, 195, 224, 237, 259, 269, 313, 314, 318, 330

Chelon labrosus (Risso, 1826) – 13, 24, 26, 32, 52, 55, 58, 62, 80, 93, 96, 98, 100, 101, 107, 110, 132, 134, 160, 178, 195, 203, 222, 224, 225, 234, 237, 259, 269, 278, 318, 313

Ciliata mustela (Linnaeus, 1758) – 13, 32, 54, 66, 93, 98, 100, 108, 150, 178, 195, 204, 269, 300, 318,

Conger conger ([Artedi, 1738] Linnaeus, 1758) – 12, 13, 32, 47, 66, 93, 98, 101, 108, 166, 178, 195, 197, 203, 204, 224, 225, 237, 247, 291, 294, 313, 318, 332

- Dicentrarchus labrax* (Linnaeus, 1758) – 13, 17, 32, 55, 62, 65, 76, 84, 88, 91, 93, 94, 95, 96, 98, 100, 101, 108, 110, 132, 134, 137, 160, 173, 177, 178, 182, 189, 195, 197, 203, 204, 215, 224, 225, 237, 259, 269, 280, 278, 294, 310, 313, 318, 329, 330
- Dicentrarchus punctatus* (Bloch, 1792) – 17, 44, 55, 62, 98, 101, 108, 125, 189, 224, 225, 237, 259, 310
- Diplodus annularis* (Linnaeus, 1758) – 6, 13, 17, 24, 26, 45, 46, 55, 62, 66, 80, 88, 93, 98, 100, 101, 108, 110, 129, 197, 204, 224, 225, 237, 251, 259, 269, 271, 278, 294
- Diplodus bellottii* (Steindachner, 1882) – 45, 55, 62, 66, 80, 98, 100, 101, 129, 203, 204, 224, 225, 237, 240, 271, 278, 294, 313
- Diplodus cervinus* (Lowe, 1841) – 13, 24, 26, 45, 93, 101, 108, 110, 129, 204, 224, 252, 253
- Diplodus puntazzo* (Cetti, 1777) – 24, 26, 44, 45, 62, 73, 80, 88, 93, 101, 107, 108, 110, 129, 204, 216, 224, 225, 237, 294
- Diplodus sargus* (Linnaeus, 1758) – 11, 12, 13, 17, 24, 25, 26, 44, 45, 55, 62, 66, 80, 88, 91, 96, 98, 100, 101, 107, 108, 110, 129, 160, 178, 195, 203, 204, 225, 237, 259, 269, 278, 294, 313, 318
- Diplodus vulgaris* (E. Geoffrey Saint-Hilaire, 1817) – 12, 13, 17, 24, 26, 32, 44, 45, 55, 62, 66, 80, 88, 93, 96, 98, 100, 101, 107, 108, 110, 129, 134, 160, 178, 195, 203, 204, 224, 225, 237, 259, 269, 278, 283, 294, 318, 313
- Echiichthys vipera* (Cuvier, 1829) – 13, 44, 93, 96, 98, 100, 110, 178, 195, 197, 237, 259, 269, 280, 312, 313, 318
- Engraulis encrasicolus* (Linnaeus, 1758) – 13, 17, 24, 26, 32, 55, 62, 66, 88, 93, 96, 98, 100, 101, 107, 108, 110, 160, 178, 195, 197, 224, 225, 231, 237, 259, 265, 269, 278, 280, 294, 318, 334, 335
- Gasterosteus aculeatus* (Linnaeus, 1758) – 9, 32, 39, 71, 91, 102, 132, 176, 178, 187, 195, 219, 234, 256, 269, 279, 285, 330, 313
- Gobius niger* Linnaeus, 1758 – 12, 13, 24, 25, 26, 32, 54, 55, 62, 66, 80, 88, 93, 96, 98, 100, 101, 106, 107, 108, 110, 131, 160, 178, 195, 197, 203, 220, 224, 234, 237, 259, 269, 271, 278, 294, 321, 313
- Gobius paganellus* Linnaeus, 1758 – 12, 13, 16, 24, 26, 32, 35, 61, 62, 66, 80, 93, 96, 98, 107, 108, 110, 171, 204, 213, 219, 220, 221, 237, 269, 294, 313
- Halobatrachus didactylus* (Bloch & Schneider, 1801) – 3, 13, 24, 25, 26, 55, 62, 66, 88, 93, 98, 101, 107, 108, 120, 203, 205, 224, 237, 271, 277, 278, 282, 294, 313
- Hippocampus hippocampus* (Linnaeus, 1758) – 13, 62, 63, 66, 80, 93, 96, 98, 101, 108, 110, 111, 145, 160, 178, 204, 205, 211, 224, 237, 259, 278, 294
- Hippocampus ramulosus* Leach, 1814 – 13, 24, 26, 93, 97, 98, 100, 101, 107, 108, 110, 111, 145, 178, 195, 197, 204, 211, 224, 237, 269
- Lampetra fluviatilis* (Linnaeus, 1758) – 8, 56, 71, 98, 272, 274, 325
- Liza aurata* (Risso, 1810) – 13, 17, 24, 25, 26, 32, 52, 53, 55, 58, 62, 66, 80, 88, 93, 96, 98, 100, 101, 107, 108, 110, 160, 166, 167, 178, 195, 224, 237, 259, 269, 278, 304, 313,

- Liza ramado* (Risso, 1826) – 13, 17, 24, 25, 26, 32, 52, 55, 58, 66, 93, 96, 98, 100, 101, 107, 108, 110, 134, 160, 163, 178, 195, 197, 203, 222, 224, 237, 259, 269, 274, 313, 318
- Liza saliens* (Risso, 1810) – 13, 24, 26, 37, 44, 52, 55, 58, 62, 80, 95, 96, 101, 110, 149, 160, 197, 224, 237, 259, 269, 278
- Microchirus azevia* (Capello, 1867) – 13, 19, 20, 48, 66, 93, 96, 97, 98, 108, 110, 160, 179, 203, 204, 237, 260, 261, 294
- Monochirus hispidus* Rafinesque, 1814 – 17, 62, 66, 80, 81, 97, 105, 108, 110, 204, 224, 260, 261, 278, 294
- Mugil cephalus* Linnaeus, 1758 – 13, 17, 24, 25, 26, 32, 52, 55, 58, 66, 91, 93, 96, 98, 101, 107, 108, 110, 130, 132, 160, 163, 178, 195, 224, 225, 237, 256, 259, 269, 273, 288, 305, 306, 313
- Mullus surmuletus* Linnaeus, 1758 – 13, 17, 24, 25, 26, 32, 44, 55, 62, 66, 76, 80, 88, 93, 96, 98, 100, 101, 107, 108, 110, 160, 164, 168, 178, 195, 204, 224, 225, 237, 259, 278, 294, 313, 318
- Nerophis lumbriciformis* (Jenyns, 1835) – 13, 63, 80, 93, 108, 110, 111, 178, 194, 195, 197, 201, 203, 204, 226, 294
- Pagellus bogaraveo* (Brünnich, 1768) – 13, 17, 45, 62, 64, 80, 93, 96, 98, 108, 110, 160, 188, 204, 217, 269, 294, 313
- Pagrus pagrus* (Linnaeus, 1758) – 13, 45, 46, 66, 93, 97, 98, 104, 108, 110, 156, 178, 185, 195, 197, 199, 208, 218, 237, 238, 284
- Parablennius gattorugine* (Linnaeus, 1758) – 13, 44, 62, 63, 66, 80, 93, 96, 98, 108, 110, 154, 160, 195, 197, 203, 234, 237, 269, 294, 336
- Parablennius pilicornis* (Cuvier, 1829) – 13, 26, 42, 62, 63, 80, 93, 108, 109, 110, 154, 178, 237, 241, 294, 296, 318, 336
- Petromyzon marinus* Linnaeus, 1758 – 4, 14, 71, 91, 132, 178, 269, 325
- Plathichthys flesus* (Linnaeus, 1758) – 2, 13, 17, 30, 32, 66, 69, 71, 91, 93, 98, 100, 104, 108, 110, 132, 151, 178, 195, 197, 234, 243, 274, 280, 204, 259, 269, 313, 318, 330
- Pomatoschistus marmoratus* (Risso, 1810) – 13, 24, 26, 93, 107, 108, 110, 213, 220, 259, 270
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- Pomatoschistus minutus* (Pallas, 1770) – 13, 24, 27, 32, 55, 62, 65, 66, 80, 91, 93, 96, 98, 100, 106, 110, 132, 155, 160, 178, 195, 196, 197, 220, 237, 259, 265, 269, 274, 278, 294, 313, 318, 330
- Pomatoschistus pictus* (Malm, 1865) – 13, 24, 26, 32, 62, 80, 93, 96, 98, 106, 107, 108, 110, 160, 178, 195, 197, 220, 234, 294
- Raja undulata* Lacepède, 1802 – 13, 59, 62, 63, 66, 85, 89, 93, 98, 103, 108, 294, 297, 298, 313, 318
- Salmo salar* Linnaeus, 1758 – 71, 91, 132, 236, 272, 301
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- Sardina pilchardus* (Walbaum, 1792) – 13, 15, 17, 24, 25, 26, 32, 55, 62, 66, 80, 93, 96, 98, 101, 107, 108, 110, 160, 161, 171, 195, 203, 212, 224, 225, 237, 259, 269, 278, 280, 294, 313, 318, 326, 333
- Sarpa salpa* (Linnaeus, 1758) – 13, 24, 25, 26, 44, 45, 62, 66, 69, 80, 88, 93, 96, 98, 100, 101, 107, 108, 110, 165, 194, 204, 224, 225, 237, 278, 322
- Scophthalmus rhombus* (Linnaeus, 1758) – 13, 18, 24, 26, 32, 62, 70, 74, 91, 93, 96, 98, 100, 101, 107, 108, 110, 160, 175, 178, 195, 204, 224, 232, 234, 237, 259, 269, 278, 286, 318, 313, 330
- Scorpaena notata* Rafinesque, 1810 – 13, 44, 55, 62, 66, 80, 88, 93, 101, 108, 170, 224, 225, 228, 230, 233, 237, 284, 294, 313
- Serranus hepatus* (Linnaeus, 1758) – 13, 24, 26, 55, 62, 66, 80, 93, 98, 101, 103, 107, 108, 110, 192, 203, 224, 225, 237, 255, 290, 294, 311, 327
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- Solea solea* (Linnaeus, 1758) – 18, 13, 31, 55, 65, 66, 67, 69, 91, 93, 95, 96, 98, 100, 101, 108, 115, 117, 118, 134, 160, 178, 193, 195, 197, 202, 204, 223, 224, 237, 245, 259, 260, 261, 269, 280, 295, 313, 318, 331
- Sparus aurata* Linnaeus, 1758 – 13, 17, 24, 26, 32, 45, 55, 62, 66, 79, 80, 88, 93, 96, 98, 100, 101, 107, 108, 110, 129, 134, 146, 160, 178, 195, 204, 224, 225, 234, 258, 259, 264, 269, 237, 278, 287, 294, 313, 337
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- Symphodus bailloni* (Valenciennes, 1839) – 12, 13, 24, 25, 26, 32, 51, 55, 62, 66, 80, 93, 96, 98, 100, 101, 108, 110, 134, 160, 178, 195, 197, 237, 259, 263, 269, 271, 278, 284, 294, 313
- Symphodus cinereus* (Bonnaterre, 1788) – 12, 13, 24, 26, 44, 51, 62, 66, 80, 93, 101, 107, 108, 110, 237, 259, 263, 271, 294
- Syngnathus abaster* Risso, 1826 – 13, 24, 26, 32, 62, 63, 66, 80, 93, 97, 98, 101, 107, 108, 110, 111, 134, 178, 195, 211, 224, 237, 259, 265, 269, 278, 294, 308, 318
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- Syngnathus typhle* Linnaeus, 1758 – 12, 13, 24, 26, 32, 44, 49, 62, 63, 66, 75, 88, 93, 98, 101, 107, 108, 110, 111, 147, 178, 195, 197, 203, 211, 224, 225, 234, 237, 246, 259, 269, 271, 308, 313
- Trachurus trachurus* (Linnaeus, 1758) – 13, 24, 26, 32, 55, 60, 62, 66, 88, 92, 93, 98, 101, 107, 108, 110, 178, 180, 195, 197, 203, 224, 225, 237, 259, 280, 292, 293, 313, 318,

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